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# **EVALUATION OF PHYSICAL CAPTURE EFFICIENCY AND DISINFECTION CAPABILITY OF A NOVEL IODINATED FILTER MEDIUM**

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5 **of a Novel Iodinated Filter Medium**  
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## 1   **Abstract**

2           A novel filter medium has been developed that combines the use of filtration and  
3 iodine disinfection to provide protection against airborne pathogens. The physical  
4 capture efficiency and biological disinfection capability of this iodinated resin medium  
5 were evaluated. Significant physical capture efficiency (> 97%) was observed for both  
6 the iodine-treated and untreated media tested, and there was no significant difference in  
7 capture efficiency between them. The efficiency was greater than 99% in many cases.  
8 The pressure drag was less than 10% of the glass fiber HEPA filter (0.0054 in  
9 H<sub>2</sub>O/(in/min) vs. 0.065 in H<sub>2</sub>O/(in/min)). Biological disinfection by the medium was  
10 evaluated using *Micrococcus luteus* and *Escherichia coli* vegetative bacterial cells. High  
11 biological deactivation efficiency was observed (99.997%). Viable penetration through  
12 the biocidal filters was observed in only 2 of 10 experiments. A near-contact mechanism  
13 in which iodine is displaced from the triiodide complex is proposed to explain the higher  
14 biological removal efficiency compared to the physical capture efficiency exhibited by  
15 the iodinated filters. The results show that an antimicrobially augmented filter medium  
16 can provide effective protection against airborne pathogens with a significantly lower  
17 pressure drop than that imposed by conventional high-efficiency filtration systems.

18  
19   **Keywords:** Iodine, Filtration, Bioaerosol, Disinfection, Pressure Drag

## 1    **Introduction**

2            As signaled by the anthrax incident in 2001, the emerging threat of bioterrorism is  
3    a great concern for national security. Among the various pathways for an attack,  
4    aerosolization is considered to be the most effective form to disperse biological agents to  
5    the widest area in the shortest time (Kortepeter and Parker 1999). The pandemic Severe  
6    Acute Respiratory Syndrome (SARS) and avian flu also clearly indicate the grave impact  
7    of bioaerosols on public health. There are many technologies that can effectively remove  
8    bioaerosols from the air. For example, High Efficiency Particle Air (HEPA) filters are  
9    commonly used to physically remove allergens (such as pollen) from the air. The capture  
10   efficiency of filters depend on several factors including the size of the challenging  
11   aerosols, the filter fibers, the velocity of airflow through the filter, and the presence or  
12   absence of electric charge on the fibers or particles (Hinds 1999). For respiratory  
13   protection of military personnel or emergency workers, bioaerosol removal technologies  
14   must meet additional requirements to be applicable. They must not exert a high demand  
15   for consumable materials or dissipate large amounts of energy. They should be simple to  
16   install and operate, versatile, and able to function in a wide range of conditions. They  
17   should also be able to handle and not be compromised by materials commonly present in  
18   battlefields or workplaces (*e.g.*, dust).

19           Another important factor in bioaerosol protection is the viability of collected  
20   microorganisms. A significant fraction of airborne microorganisms remain viable after  
21   collection on filtration devices. There is a great potential for microbial growth on filters  
22   under favorable conditions, because of this remaining viability. Proper nutrition and  
23   moisture conditions may allow microbial growth and then subsequent re-entrainment

1 from filter media. For example, molds are able to grow on fibrous media if provided with  
2 70–80% relative humidity and atmospheric dust (Maus et al. 2000). Bacterial and mold  
3 spores collected on air filter media can survive over prolonged periods of time and pose a  
4 potential for microbial growth, especially when humidity is high ( $RH > 70\%$ ) and filters  
5 are not exposed to airflow. Abundant production and gradual release of spores into the  
6 clean-air stream of the filters is likely to occur (Maus et al. 2000). Conventional  
7 filtration, therefore, may not offer the best solution for respiratory protection according to  
8 the reasons described above. Large pressure drop can lead to breathing difficulties and  
9 subsequently reduced mobility, agility and stamina of the protected person. Air leakage  
10 also increases with filter pressure drop. The inability of conventional media to disinfect  
11 collected microorganisms also needs to be overcome.

12       The halogens iodine and chlorine are antimicrobial agents of great importance.  
13 Halogen disinfection is a form of chemical sterilization in which oxidation of cell  
14 constituents and halogenation of cell proteins occurs (Prescott et al. 2002). Iodine has  
15 been used as a disinfectant for potable and on-site water treatment, and is known for its  
16 stable chemical storage characteristics (Brion and Silverstein 1999). Iodine in the  
17 oxidation state of zero ( $I_2$ ) is not highly soluble in water but may be introduced by heat  
18 vaporization, crystal dissolution, oxidation of iodide ( $I^-$ ) ion, and release from iodine-  
19 containing resins or from the direct addition of high-strength iodine/alcohol solutions or  
20 triiodide ( $I_3^-$ ) ions (Black et al. 1968). Iodine has the ability to bind to quaternary  
21 ammonium anion exchange resins as tri- or penta-iodide complexes (Berg et al. 1964;  
22 Brion and Silverstein 1999; Chang 1958; Taylor et al. 1970). These complexes are a



1 demand type disinfectant, releasing iodine only when needed, thus allowing for a longer  
2 lifetime of product and minimizing casual exposure to iodine.

3       The use of filtration in combination with an iodinated resin has previously been  
4 adopted for water disinfection aboard spacecraft. Such a resin consists of polyiodide  
5 anions bound to quaternary ammonium fixed positive charges on a poly(styrene–  
6 divinylbenzene) copolymer anion exchange resin (Marchin et al. 1997). A stable ionic  
7 bond is formed between the iodine complex and the resin polymer, controlling the release  
8 of free iodine for disinfection of microorganisms. The bound polyiodide anions release I<sub>2</sub>  
9 into water when they come in contact with suspended microorganisms, which results in  
10 “devitalization” of the microorganism due to oxidation that hinders its cellular functions.

11       The use of such an iodinated resin product in combination with filtration has also  
12 been proposed for the removal of bioaerosols, although there is only limited research  
13 reported about the use of this product to disinfect air. According to a patent by Messier  
14 (2000), such a resin can achieve high removal efficiency of microorganisms in air under  
15 various conditions. Further evaluation of the iodinated resin product, however, is needed  
16 to characterize its removal. The objective of this study was, hence, to appraise the use of  
17 iodine in the iodine-exchanged resin for air disinfection. This was accomplished by  
18 assessing the biological removal efficiency in comparison with its physical removal  
19 efficiency. Its pressure drag was compared to that of a glass fiber filter. The effects of  
20 flow rate and filter thickness were investigated. Two types of vegetative cells were tested  
21 in addition to inorganic aerosols.

## 1    **Experimental Methodology**

2            The experiments were carried out in two phases. In Phase I, the physical removal  
3    efficiency was evaluated using ammonium fluorescein aerosols. In Phase II, two types of  
4    vegetative cells were challenged to assess the filter's biological removal efficiency. The  
5    same experimental system was used in both phases.

## 7    ***Experimental Set-up***

8            Figure 1 shows the schematic of the experimental set-up. Aerosols were  
9    generated using a six-jet Collison nebulizer (Model CN25, BGI Inc.) at a flow rate of  
10   10 Lpm for Phase I experiments. Filtered compressed air passing through a diffusion  
11   dryer was used as dilution air and was introduced into the dilution dryer chamber at twice  
12   the amount of the aerosol flow to evaporate the water content of the droplets (May 1972).  
13   An excess airflow point was used to control the flow going through the target filter at the  
14   designated level (13, 15, and 21 Lpm) while maintaining the flow rate of 28.3 Lpm  
15   required for the cascade impactor. For the bioaerosol experiments, 7 Lpm of air was used  
16   for the nebulizer, and dilution air was 13 Lpm. Additional air was introduced  
17   downstream of the dryer to maintain 15 Lpm passing through the filter. Relative  
18   humidity and temperature downstream of the dilution dryer were monitored (HX 94,  
19   Relative Humidity/Temperature Transmitters, Omega Engineering). The aerosols were  
20   then passed through the test filter (or empty filter holder in the control runs). A  
21   Magnehelic gage reading 0–10 in H<sub>2</sub>O was employed to evaluate the pressure drop across  
22   the resin/iodine filters. A reading was taken every minute for each run. Pressure drop is  
23   due to the resistance to airflow across a filter. Penetrating aerosols leaving the test filters

1 were captured and classified by particle size on a six-stage Andersen viable impactor  
2 (Model #10-820). All the flow rates were controlled by pre-calibrated rotameters.

3

#### 4 ***Test Particles***

5 In Phase I, ammonium fluorescein particles were employed to evaluate the  
6 physical capture efficiency of the filters due to the lower detection limit of fluorescence  
7 compared to gravimetric measurements. A 6.75-g/L fluorescein solution in 0.1 N NH<sub>4</sub>OH  
8 was aerosolized by the Collison nebulizer. This concentration was chosen to allow for the  
9 production of larger particles and enhanced detection. The mass median diameter of the  
10 dry fluorescein particles was calculated to be ~ 0.27 µm, based on the following equation  
11 (Hinds 1999):

$$12 \quad d_a = d_d (F_v)^{1/3} \quad (1)$$

13 where  $d_d$  is the mass median diameter of the atomized droplet (~ 3 µm; (May 1972) and  
14  $F_v$  is the volume fraction of fluorescein in the solution.

15 In Phase II, microorganisms for bioaerosol challenges were selected based on  
16 several factors. Nonpathogenic representatives of possible biological weaponry agents  
17 were employed. A commonly known airborne pathogen is *Yersinia pestis*, the organism  
18 that causes bubonic plague (Morris and Sandana 2005). As a representative for this  
19 organism, *Escherichia coli* was utilized. *E. coli* is a Gram-negative rod-shaped bacterium  
20 that ranges in size from 2 to 3 µm in length and 0.25 to 1 µm in diameter. The strain used  
21 was obtained from the Water Reclamation Facility in the Department of Environmental  
22 Engineering Sciences at the University of Florida. The samples obtained were inoculated  
23 and maintained on Difco tryptic soy agar and grown at 33°C prior to sampling.

*Micrococcus luteus* is another frequently used representative bioaerosol (Agranovski et al. 2003; Li and Lin 2001; Wake et al. 1997). *M. luteus* cells are Gram-positive, non-motile, nonsporulating, round bacteria normally found in clusters or tetrads. The individual cells are 0.9 to 1.8  $\mu\text{m}$  in diameter (Wake et al. 1997). *M. luteus* samples were obtained from the University of Florida, Department of Microbiology and Cell Sciences. Cells were inoculated on standard nutrient agar (Difco 0001) and maintained at 33 °C prior to sampling. Prior studies (Crook et al. 1998; Wake et al. 1997) have shown that Ringer's solution can successfully maintain the viability of stressed bacteria used in aerosol studies. Hence, bioaerosol suspensions for use in the Collison nebulizer were created by washing bacterial cells off agar slants using 1 mL of 25% Ringer's Solution (Fisher, S77939). The slants were agitated for 20 seconds using a standard vortex, and varying amounts of each sample were aseptically pipetted out of the slant test tubes and into Ringer's solution contained in the nebulizer reservoir for each experiment.

#### ***Test Filter***

Filter media were provided by Triosyn Corp. through the U.S. Air Force Research Laboratory. The iodinated resin is produced by thermally fusing pure iodine crystals with a quaternary anion exchange polymer under high pressure. Iodine-treated and untreated filters of 1 mm thickness were tested for physical capture efficiency and to evaluate whether differences in morphology due to the iodine treatment would affect capture efficiency. Heavier filters of 2 mm thickness were used to evaluate the effect of depth. The bioaerosol experiments were conducted using medium-depth filters with an approximate thickness of about 1.5 mm.

1           The filter media tested were 47 mm in diameter (area 17.35 cm<sup>2</sup>). A common  
2   respirator cartridge has a cross-section area of approximately 100 cm<sup>2</sup> and a nominal  
3   breathing rate is about 85 Lpm (Di Ionno and Messier 2004). Accordingly, the flow rate  
4   used for testing the 47-mm filter was scaled down to 15 Lpm to produce a similar face  
5   velocity. Two other flow rates, 13 and 21 Lpm, were used to evaluate the effects of flow  
6   velocity. Bioaerosol challenges were conducted only at 15 Lpm air flow rate across filter  
7   surfaces.

8

## 9   *Experimental Procedure*

### 10   Phase I

11           Each experiment was run for 15 minutes. This amount of time was shown to be  
12   sufficient to deliver a measurable amount of fluorescent particles for evaluation, while  
13   not causing a mound effect due to accumulation of particles under each impactor jet,  
14   which might alter the collection characteristics of the Andersen impactor. After  
15   collection, the individual stages containing fluorescein particles were rehydrated in  
16   aqueous 0.1 N NH<sub>4</sub>OH solution. They were then treated with 20 mL of methylene  
17   chloride to dissolve the grease coating on each plate and sonicated for 10 minutes.  
18   Twenty mL of 0.1 N NH<sub>4</sub>OH was added to each sample after sonication and poured into a  
19   test tube. Due to the immiscibility of methylene chloride and ammonium hydroxide, a  
20   fluorescein–ammonium hydroxide solution separated to the top of each sample while the  
21   methylene chloride, containing the grease, sank to the bottom. The top layer of each  
22   sample was then pipetted into a quartz cuvette for analysis (Vanderpool et al. 1987).

Mass concentrations of fluorescein solution from each stage were measured using a Sequoia–Turner 112 Digital Filter Fluorometer (G.K. Turner Associates). A calibration of concentration vs. fluorometer reading was performed using samples of known fluorescein concentration in 0.1 N NH<sub>4</sub>OH for fluorometer range settings of 1X, 3X, 10X, and 30X. Fluorometer readings less than 30 and higher than 100 were considered to be less reliable and therefore were not used in this study. Accuracy was maintained by selecting range settings that best suited the concentration of the solution being analyzed.

Several factors may skew the results of fluorometric analysis, including fluorescence given off by reagents, filter substances reacting with chemicals, or flaws within different cuvettes used in analysis. For this reason, several background tests for the various filter media were conducted. The results indicate that there was no interference with fluorescence analysis by the glass fiber and untreated filters. Variations in the optical properties of each cuvette were also shown to have negligible effect on data reproducibility. A maximum standard deviation of only 0.8% was observed at 500 µg/L based on 36 fluorometric readings of different cuvettes oriented at different directions and at differing concentrations. However, a negative interference was observed for the iodine-treated filters. The interference may be attributed to chemical reactions occurring between the iodine and fluorescein in the rehydrated filter solution, and it can result in a lower apparent concentration. A concentration-based interference curve was therefore established to adjust the values measured.

## Phase II

To measure the inlet concentration of bioaerosols entering the test system vs. those captured by filtration, two impactors were used in parallel. One impactor contained no filter upstream, whilst the other contained an iodine-treated filter upstream. The inlet concentrations were measured for the first and last five minutes of each experiment using Petri dishes on all six stages of the impactor. Due to the very low outlet concentration or penetrating concentration, the bioaerosol was collected on only one Petri dish on stage six (last stage). Every 20 minutes throughout the 2 h run, the outlet Petri dishes were changed out to prevent desiccation of the agar surface. After each plate was removed from the impactors, it was labeled and placed in an incubator at 33 °C for 24–36 h. It was noted that growth of colonies on the surface of the impaction zone was in the same pattern as in the Phase I experiments. The optical count of the number of positive holes in each sample was corrected for the number of colony-forming units (cfu) that impacted onto an agar surface following the positive hole method (Thermo Electron Corp., 2003).

## *Statistical Analysis of Data*

Three controls (with no filters upstream) and three experimental runs (with filters upstream) were conducted separately for each flow rate tested. The least squares method was used to correlate the data (Vining 1998). This method considers the data points assessed during the experiments and finds the best straight-line equation to represent all combinations of the data. The  $x$ -axis represents the control mass fraction measured for each stage for this analysis, and the  $y$ -axis represents the difference between the control mass fraction and the penetration mass fraction for each stage. The efficiencies reported

were the slopes calculated based on the least squares methods, and the error ranges reported were the standard errors for the y estimates.

#### ***Filter Morphology Analysis***

Microscopic images of the filters before and after sampling were taken using scanning electron microscopy (SEM) to evaluate any differences in morphology due to treatment or use of the filters. Elemental analysis of objects magnified with SEM was performed using energy dispersive X-ray (EDX) analysis.

### **Results and Discussions**

#### ***Morphology Analysis of Filter Media***

Iodine-untreated filters were analyzed via SEM prior to experimentation (Figures 2a and 2b). A dense woven structure of long fibers of different thickness was observed by visual analysis. Iodine-treated filters were also analyzed via SEM prior to experimentation (Figures 2c and 2d). Small dark flecks were observed around the outer perimeter of the filter surfaces. EDX analysis was used upon magnification to determine the elemental composition of fibers (Figure 2c) and black specs (Figure 2d) seen in the SEM photographs. The EDX analysis of the small flecks indicated the presence of iodine (Figure 3a). Similar results were observed for the EDX taken of the filter fibers, demonstrating the presence of iodine on the fiber surface.

#### ***Phase I – Physical Capture***



1           The mass size distribution of particles produced in the system reaching the point  
2 of filtration was determined during control runs using no filter upstream of the impactor.  
3 Figure 4 shows the size distribution at 15 Lpm as an example. The shape of the  
4 distribution of other flow rates had a similar pattern, while the total mass increased as the  
5 flow rate increased. The majority of the fluorescent particles were collected on the  
6 downstream filter stage ( $<0.65\ \mu\text{m}$ ) and 5th and 6th stages ( $2.1\text{--}1.1$  and  $1.1\text{--}0.65\ \mu\text{m}$ ,  
7 respectively) for each flow rate. The data for stages 1 to 4, however, were not used due  
8 to the low concentrations of detectable particles. The mean corresponding total mass  
9 concentrations were  $1.33\times 10^7$ ,  $2.14\times 10^7$ , and  $4.03\times 10^7\ \mu\text{g}/\text{m}^3$  for 13, 15, and 21 Lpm,  
10 respectively.

11           Table 1 summarizes the physical capture efficiency at each stage of the impactor  
12 for iodine-treated and untreated filters at different flow rates. Figure 4b shows, as an  
13 example, the mass size distribution downstream from an iodinated resin filter. As shown,  
14 significant capture (greater than 97%) was observed for both the iodine-treated and  
15 untreated filters tested for stages 5 and 6 and for the downstream filter (DF). Filters were  
16 visually analyzed after sampling and rehydration. It was observed that some amount of  
17 fluorescein was not captured by the upstream iodinated resin filter. No fluorescence was  
18 observed on the downstream filters when similar experiments were performed using glass  
19 fiber filters upstream, indicating that the glass fiber filters have higher removal efficiency  
20 for fluorescence particles than the iodinated resin filters do. Rehydrated solutions in test  
21 tubes were also visually analyzed prior to pipetting into cuvettes. No visible fluorescence  
22 was observed when glass fiber filters were located upstream, whereas visible  
23 fluorescence could be seen from the iodinated resin samples.

Removal efficiency greater than 99.8% was observed for the 1.1–2.1  $\mu\text{m}$  particles for all flow rates tested. The efficiency decreased to less than 99% for particles smaller than 0.65  $\mu\text{m}$  and to 96.84% for the iodine-treated filters tested at 15 Lpm. Impaction is the most important mechanism for the larger particles (1.1–2.1  $\mu\text{m}$ ). Hence, increasing the flow velocity (21 Lpm vs. 13 Lpm) resulted in higher efficiency. Conversely, the dominant capture mechanism for smaller particles (less than 0.65  $\mu\text{m}$ ) is diffusion. A higher flow rate resulted in shorter retention time for diffusion and consequently lower collection efficiency. The highest capture efficiency was generally observed in the 1.1–2.1  $\mu\text{m}$  range, and efficiency decreased as the particle size decreased. Both treated and untreated filters appeared to perform similarly based on the data, which showed no significant difference. The thicker filters (2 mm) appeared to perform better than the regular iodine-treated filters at 15 Lpm. This was expected, due to the increased possibility for impaction and longer retention time leading to greater diffusion of particles to fiber surfaces. However, the improvement was not huge because the thin filters (1 mm) already had exceptional capture efficiency.

Pressure drop across the filters was recorded. The system was tested with and without the use of the aerosolized particles to determine how particle accumulation on the filter affected pressure drop. A pressure drop of 0.2 in  $\text{H}_2\text{O}$  was observed when no filter was inserted into the filter holder, which was attributed to the wire mesh backing that was placed behind the filter for support. The initial pressure drop when the filter (1 mm thickness) was placed was 1.8 in  $\text{H}_2\text{O}$  at 15 Lpm and 2.3 in  $\text{H}_2\text{O}$  for 21 Lpm. It then increased as the particles flowed through the testing filters and were subsequently captured.

1 Pressure drag ( $S$ ) is a measure of the filter's aerodynamic resistance to air flow  
2 and can be calculated by dividing pressure drop across a filter ( $\Delta p$ ) by filtration velocity  
3 ( $U$ ) as (Noll 1999):

$$4 \quad S = \frac{\Delta p}{U} \quad (2)$$

5 It is worthwhile noting that the initial pressure drag of the iodinated filters was  
6 significantly less than the associated pressure drag of the glass fiber filters tested,  
7 0.0054 in  $\text{H}_2\text{O}/(\text{in}/\text{min})$  vs. 0.065 in  $\text{H}_2\text{O}/(\text{in}/\text{min})$ . Lower filter drag is associated with  
8 less-labored respiration, which is beneficial when the mobility of the protected person is  
9 critical. Another useful criterion for comparing different types of filters is filter quality,  
10  $q_F$  (Hinds 1999),

$$11 \quad q_F = \frac{\ln(1/P)}{\Delta p} \quad (3)$$

12 where  $P$  is the aerosol penetration through the filter. The iodinated filter's filter quality  
13 was  $9.2 \text{ kPa}^{-1}$ , which is higher than the glass fiber's value, 1.8. It again supports that the  
14 iodinated filter exhibits better quality.

15

## 16 ***Phase II — Biological Disinfection***

17 Two types of bacteria were evaluated for the bioaerosol challenges, *E. coli* and *M.*  
18 *luteus*. The size distributions of colony-forming units detected at the inlet are shown in  
19 Figure 5. As shown, the majority of the bioaerosol particles generated were detected in  
20 the  $1.1\text{--}2.1 \mu\text{m}$  range, which agrees with the nominal sizes reported in the literature. The  
21 overall disinfection efficiency was  $99.997 \pm 0.004\%$  for *M. luteus* and  $99.998 \pm 0.005\%$   
22 for *E. coli*. It should be noted that penetration was detected in only two of 10

1 experiments. These results indicate that the iodinated resin filters cause close to a 5-log  
2 average removal of bioaerosols when tested using the methods discussed in this study.  
3 The similar efficiencies for both species also suggest that the mechanism works for both  
4 Gram-positive and -negative species.

5 In addition, disinfection removal efficiency was approximately 2 logs higher than  
6 the physical efficiency (99.8% physical capture efficiency at 15 Lpm, Table 1). This  
7 removal efficiency increase may be explained by reaction with gaseous iodine present in  
8 the filtration system. However, measurements of  $I_2$  vapor downstream applications of  
9 similar iodinated polymer materials showed low iodine vapor concentration ( $< 0.2$

10  $mg/m^3$ ; OSHA TLV is  $1 mg/m^3$ ) (Di Ionno and Messier 2004; Di Ionno et al. 2001).

11 Another possible explanation is near-contact transfer of  $I_2$  as the bacteria pass near the  $I_3^-$   
12 complex. Figure 6 displays the concept. Bacteria are almost universally anionic at their  
13 surface. When they fly in proximity to the  $I_3^-$  complex on the polymer surface, the charge  
14 on the microbe surface displaces the  $I^-$  ion and captures the  $I_2$  molecule. The  $I_2$  molecule  
15 then reacts with an iodlatable group on the microbe. Thus, disinfection can occur near  
16 but without direct contact with filter surface. Confirmation of this hypothesis, however,  
17 requires further investigation.

18 The initial pressure drag of the 1.5-mm thickness filter was 0.006 in  
19  $H_2O/(in/min)$ , which is still less than 10% of that of a glass fiber HEPA filter. The filter  
20 quality based on the biological removal was  $19.9 kPa^{-1}$ . The high biological disinfection  
21 efficiency, low pressure drag and high filter quality together demonstrate that the  
22 iodinated filter offers a superior alternative to conventional HEPA filtration for removing  
23 biological agents.

1 Combinations of SEM with EDX were used to evaluate filters after experimental  
2 runs with *M. luteus* suspended in 25% Ringer's solution. Figures 2e and 2f display their  
3 images. Particles of various size and shape were observed in contact with the filter  
4 fibers. The filter fibers appear to be coated in a white substance. Also notable is the  
5 absence of the large, dark specs observed in the filters prior to experimentation. The  
6 EDX analysis (Figure 3b) verified that the particles were primarily composed of sodium  
7 and chlorine, indicative of the Ringer's solution used in the experiment. The coating is  
8 expected to have no influence on the viability of penetrating microorganisms. It is,  
9 however, not clear what impact—if any—the coating had on the viability of those cells  
10 collected on the filter. It should be noted that the Ringer's solution is necessary to keep  
11 the microorganisms alive during the experiments, since pure water lyses the cells. Thus,  
12 it is unlikely that the undesired presence of the materials can be avoided in the testing.

13 The viability of those cells captured on the filter is also of interest. Unfortunately,  
14 this could not be verified because the retrieval of those collected microorganisms requires  
15 the use of water. Iodine released into the water can easily disinfect those cells as reported  
16 in numerous studies (Berg et al. 1964; Black et al. 1968; Marchin et al. 1997), thus  
17 yielding no viable cells for verification. Nevertheless, from the low level of penetration,  
18 it can be expected that the possibility of survival on the filter is rather low.

19 It is well known that many microorganisms aggregate. There are concerns  
20 whether aggregation would provide shielding protection to those hiding in the core. This  
21 is likely not to be an issue for vegetative cells in the iodinated resin filter system. Large  
22 aggregates of vegetative cells are at least a few micrometers, and hence they will be very  
23 efficiently captured by the physical mechanism. Gas-phase or captured iodine can also

penetrate the aggregates to reach the bacteria inside. Thus, aggregated vegetative cells pose no greater threat when such a filtration system is employed. The same statement, however, may not be extended to virus particles. Virus aggregates may still well be in the submicrometer range, where the collection efficiency is lower.

## Conclusions

In this study, the physical capture and biological disinfection efficiency of a novel biocidal filter medium were evaluated. Significant capture (greater than 97%) by the filters was observed for a wide particle size range. In most cases the efficiency was greater than 99%. Efficiency was the highest for larger particles (1.1–2.1  $\mu\text{m}$ ), and it increased slightly as flow velocity increased, since impaction is the main collection mechanism for this size range. Efficiency was the lowest for smaller particles (less than 0.65  $\mu\text{m}$ , collected in the downstream filter), and it decreased as flow velocity increased since diffusion is more important for smaller particles. No discernible difference in the physical capture efficiency between the iodine-treated and untreated filters suggests no mechanical impact of the treatment. The iodinated resin filters were not as efficient at physical removal of aerosols as the glass fiber HEPA filter; nonetheless, the pressure drag was 90% less and the filter quality for biological removal was higher. Enhanced capture efficiency can be easily achieved by using a thicker filter while still being in the acceptable range of pressure drop.

Two types of microorganisms were used in the bioaerosol challenge experiments, *M. luteus* and *E. coli*, which were dominantly in the 1.1–2.1  $\mu\text{m}$  aerodynamic range. There was no difference in removal efficiency between these two species representing Gram-positive and -negative bacteria; the average efficiency was approximately

1 99.997%, or close to 5-log removal of bioaerosols. Indeed, only two of 10 experiments  
2 performed had detectable penetration through the iodine-treated filters. This efficiency  
3 showed much greater biological removal than was physically established. This may be  
4 indicative of the presence of gaseous iodine within the filtration system or downstream of  
5 the filters or, as we propose, of a direct mechanism involving capture of  $I_2$  from the  
6 medium. Whether the bacterial cells collected on the filter were viable could not be  
7 verified due to the well known disinfection capability of iodine in water. Nonetheless,  
8 the low penetration fraction of viable bioaerosols implies an extremely low survival  
9 fraction, if any. The high biological disinfection capacity combined with the low  
10 pressure drag and high filter quality demonstrates the novel filter medium to be a superior  
11 alternative to conventional filtration for the removal of micrometer bioaerosols. Its  
12 application to much smaller virus aggregates, however, needs to be investigated. The  
13 effective lifetime should also be determined.

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17  
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Table 1 Physical Capture Efficiency of Iodinated Resin Filters per Stage for 1 mm thick filters

Sample, Flow rate	Stage# (Size range, $\mu\text{m}$ )				Total (2.1–0.03)
	5 (2.1–1.1)	6 (1.1–0.65)	DF (0.65–0.03)	DF (0.65–0.03)	
Treated @ 13 Lpm	99.86 $\pm$ 0.0111	99.96 $\pm$ 0.0292	99.22 $\pm$ 0.167	99.34 $\pm$ 0.118	
Untreated @ 13 Lpm	99.91 $\pm$ 0.00447	99.79 $\pm$ 0.0263	99.12 $\pm$ 0.184	99.62 $\pm$ 0.830	
Treated @ 15 Lpm	99.81 $\pm$ 0.00509	99.21 $\pm$ 0.0464	96.84 $\pm$ 0.0634	97.32 $\pm$ 0.381	
Untreated @ 15 Lpm	99.89 $\pm$ 0.0328	99.87 $\pm$ 0.0234	99.38 $\pm$ 0.101	99.43 $\pm$ 0.0826	
Treated @ 21 Lpm	99.99 $\pm$ 0.000308	99.51 $\pm$ 0.0199	98.85 $\pm$ 0.149	99.01 $\pm$ 0.0877	
Untreated @ 21 Lpm	99.93 $\pm$ 0.0320	99.53 $\pm$ 0.0308	98.95 $\pm$ 0.263	99.07 $\pm$ 0.191	
Thick* Treated @ 15 Lpm	99.94 $\pm$ 0.00276	99.97 $\pm$ 0.00191	99.50 $\pm$ 0.126	99.56 $\pm$ 0.0904	

\* 2 mm



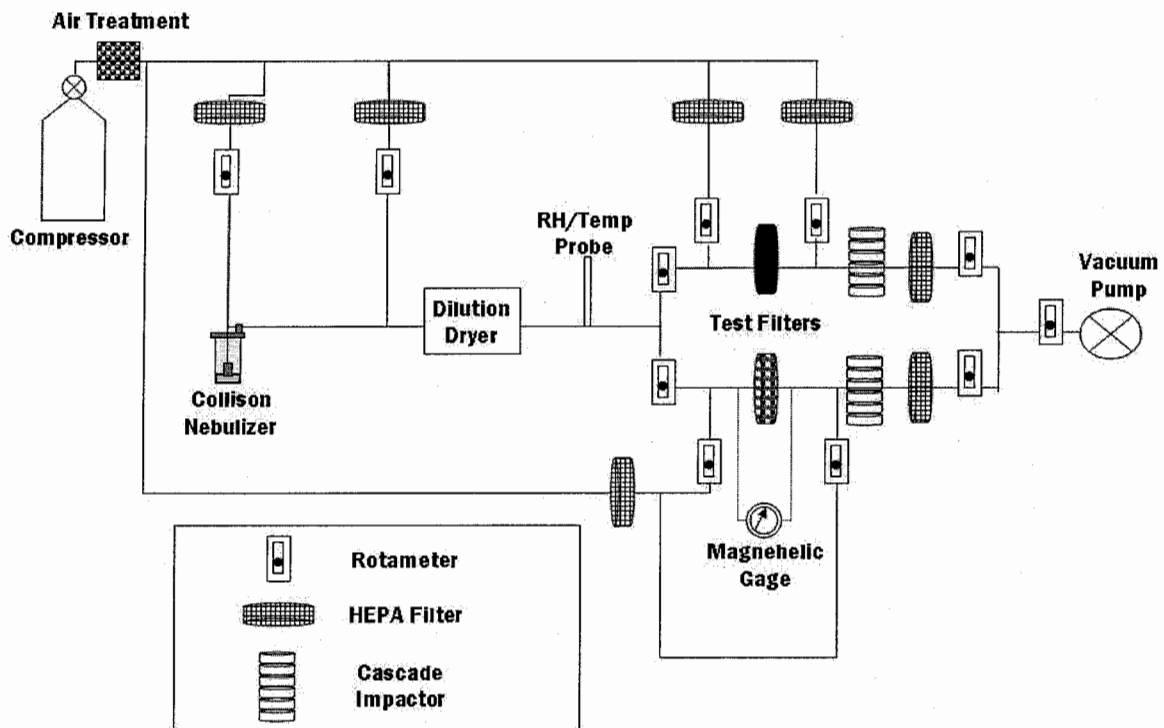


Figure 1 Experimental Set-Up

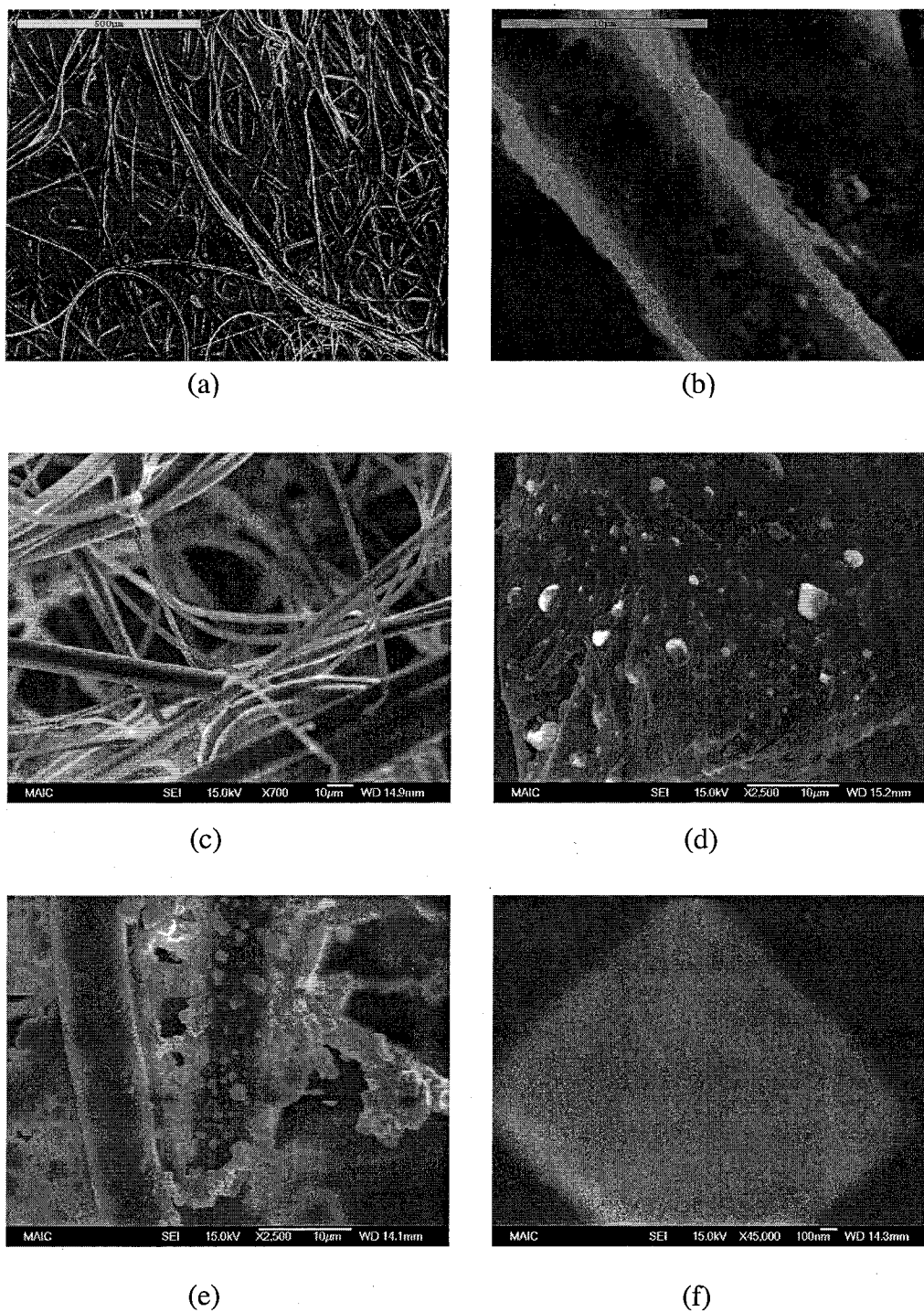
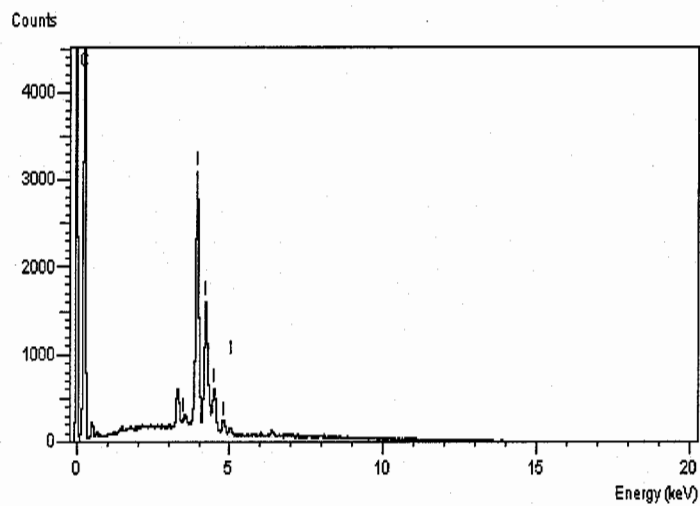
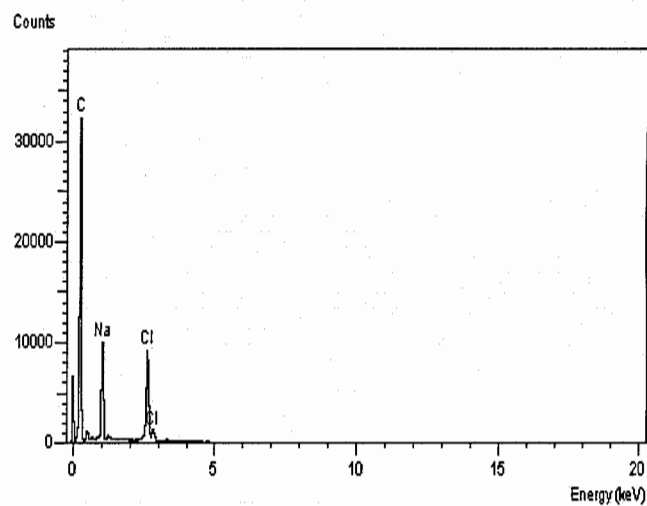


Figure 2 SEM images of untreated filters (a) 5100X, (b) 54000X; SEM images of fresh iodine treated filters: (c) fibers at 700x, and (d) enlarged fleck at 2500X; SEM images of used filters: (e) 2500X, and (f) enlarged particle at 45000X



(a)



(b)

Figure 3 EDX spectrum: (a) enlarged fleck on fresh iodine-treated filter (Figure 2d), (b) enlarged particle on used filter (Figure 7(f))

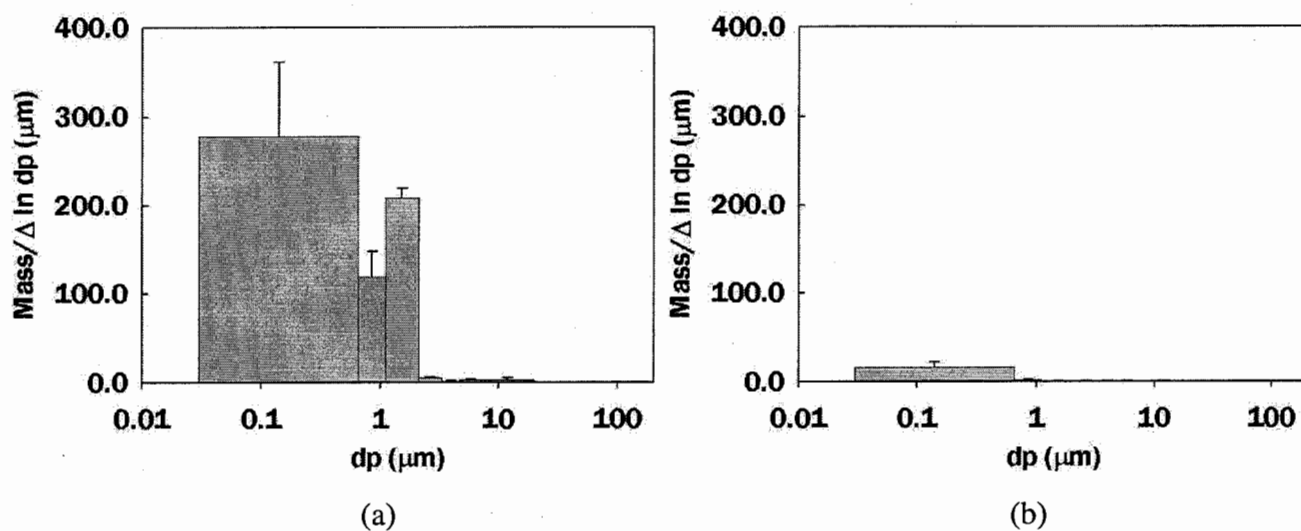


Figure 4 Mass size distribution of ammonium fluorescein particles for Phase 1 at 15 Lpm: (a) control experiments, (b) with an iodinated filter upstream

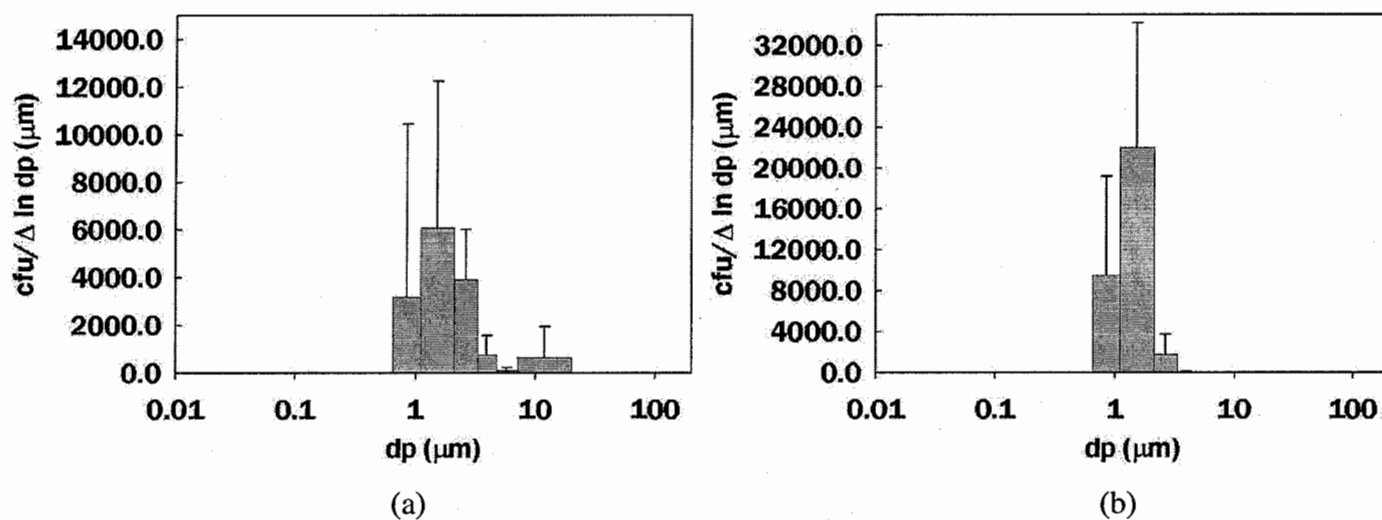


Figure 5 Size distribution of bioaerosols generated at 15 Lpm for (a) *M. luteus*, and (b) *E. coli*

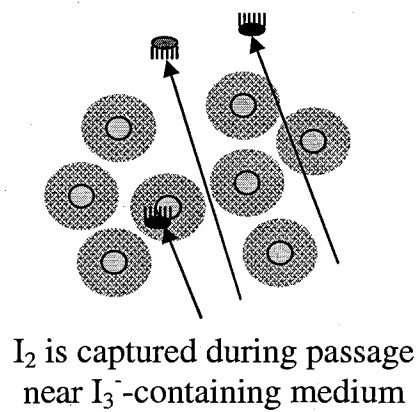
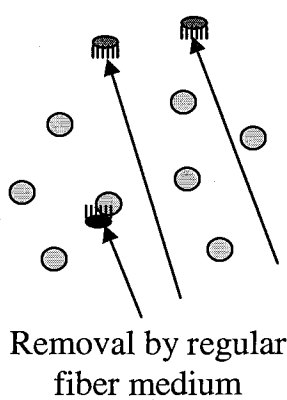
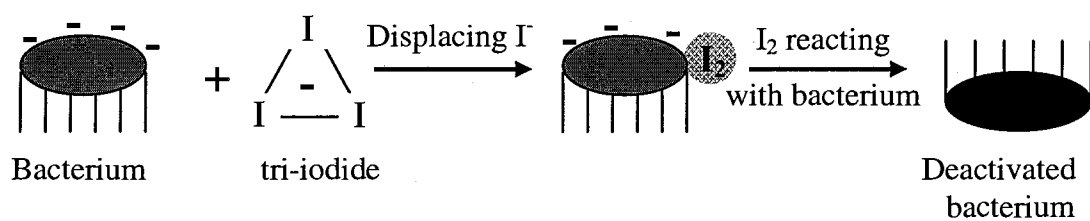


Figure 6 Conceptual schematic of disinfection near the iodinated polymer

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